Immobilization of ⁹⁹Tc (Re) using Iron-Phosphate Glasses

Jong Heo^{a*}, Kai Xu^a, Wooyong Um^{a,b}, Pavel Hrma^{a,b}

^a Division of Advanced Nuclear Engineering and Department of Material Science and Engineering,

Pohang Univ. of Science and Technology (POSTECH), Pohang 790784, S. Korea

^b Pacific Northwest National Laboratory (PNNL), P.O. Box 999, Richland, Washington 99354, USA

**Corresponding author: jheo@postech.ac.kr*

1. Introduction

Technetium-99 (99Tc) is a fission product artificially generated during the irradiation of ²³⁵U for commercial power production or ²³⁹Pu for nuclear weapons. Under oxidizing conditions, the dominant species of Tc, the pertechnetate anion (TcO_4) , is highly soluble in ground water and thus easily transports through the geologic systems. In addition, because of its high fission yield (~6 %) and long half-life (2.1×10^5) yr), immobilization of ⁹⁹Tc has been investigated for decades [1,2]. Several waste forms such as metallic alloys, sintered titanate ceramics and chemically bonded phosphate ceramics have been proposed to encapsulate 99Tc. They have not yet been realized in the industrial-scale, mostly either due to the high volatilization of ⁹⁹Tc during high temperature process $(>1300^{\circ}C)$, or the low ⁹⁹Tc loading [3-5].

Iron-phosphate (FeP) glasses have been developed as alternative waste forms because of their chemical durability equivalent to borosilicate glasses [6]. Additionally, vitrification of radioactive waste by FeP glasses can be done at a relatively low temperature (~1000°C) and the low-temperature process can reduce the volatilization of ⁹⁹Tc significantly. Thus, this work reports the immobilization of ⁹⁹Tc by FeP glasses using rhenium (Re) as a surrogate. We also examine the chemical durability of Re-containing FeP glasses using product consistency test (PCT). Experimental results reveal that FeP glass can become a promising candidate for immobilizing ⁹⁹Tc.

2. Experimental Methods

FeP waste forms were prepared by a two-step procedure. First, base glasses with a nominal composition of $42P_2O_5$ - $25Na_2O$ - $5Al_2O_3$ - $10CaF_2$ - $18Fe_3O_4$ (mol %) were melted at $1050^{\circ}C$ for 30 min. Melt was then quenched and crushed into powder. Then, the powdered glass was mixed with different amounts of KReO₄ (1.5, 2, 3, 4 and 6 wt. %) and remelted at $1000^{\circ}C$ for 10 min before pouring into a brass mold.

Crystalline phases in the waste forms were identified with X-ray diffraction (XRD). The retention of Re was determined with the inductively coupled plasma mass spectrometry (ICP-MS). Distribution of Re in the glass was ascertained with the energy dispersive X-ray spectroscopy (EDX). Chemical durability of Re-containing phases was examined with 7-day PCT following ASTM C 1258-02. Elements in the solution were analyzed with ICP-MS. Normalized elemental mass release, r_i (g/m²), was calculated as:

$$r_i = \frac{C_i}{f_i(A/V)} \tag{1}$$

where C_i is the concentration of element *i* in the solution (µg/ml or g/m³), f_i is the mass fraction of element *i* in the glass (unitless), and A/V is the ratio of the sample surface area to volume of leachate (m⁻¹); $A/V = 1790 \text{ m}^{-1}$.

3. Results and Discussion

FeP glass with 1.5 wt. % of KReO₄ addition contained tiny crystals only visible under optical microscopy, while other glasses had <10 wt. % of crystals in the form of spherical inclusions. These crystals were removed from the glass and analyzed with XRD as shown in Fig. 1. The main crystalline phase was $Fe_2Na_3(PO_4)_3$ when < 3 wt. % of KReO₄ was added. With > 3 wt. % KReO₄ added, NaReO₄ and KReO₄ crystals settled down in the glass. This result suggests that the initial addition of KReO₄ in FeP base glasses should be < 3 wt. %.



Fig. 1. XRD patterns of crystals in samples with 2, 3, 4 and 6 wt. % KReO₄ addition.

Actual concentration of Re retained in waste forms analyzed by ICP-MS was as high as ~1.1 wt. % when more than 4 wt. % of KReO₄ was added into FeP glasses (Table I). As Table I shows, Re volatilization can be reduced to ~50 % and it is considerably smaller than those (~90 %) found from normal borosilicate glasses [7]. EDX mapping of Re in FeP waste forms shows Re distribution was uniform and no Re-rich phase was detected (Fig. 2).

Table I: Re retention in FeP glass with different KReO₄ additions.

KReO ₄	Re	Re	Volatilization
additioin	addition	Retained	(%)
(wt. %)	(wt. %)	(wt. %)	
1.5	0.97	0.50	48
2	1.29	0.65	50
3	1.93	0.88	54
4*	2.57	1.12	56
6*	3.86	1.13	71

*: White crystals of NaReO₄ or KReO₄ persisted in the glass.



Fig. 2. EDX mapping of Re in FeP glass with different KReO₄ addition.

The normalized 7-day PCT elemental releases, r_i , of Re, Na, P and Al are listed in Table II. Values of r_{Re} were $< 10^{-2}$ g/m² and are comparable with values measured from other waste forms [3, 5, 7]. Releases of other elements (Na, P and Al) indicate that the FeP glass used has a high chemical durability.

Table II: Normalized 7-day PCT elemental mass releases (g/m²) from FeP glass with different KReO₄ additions

KReO ₄ (wt. %)	r _{Re}	<i>r_{Na}</i>	r _P	r_{Al}
1.5	8.4×10 ⁻²	0.15	4.3×10 ⁻²	1.9×10 ⁻²
2	6.3×10 ⁻²	0.19	3.8×10 ⁻²	5.8×10 ⁻³
3	6.4×10 ⁻²	0.17	5.1×10 ⁻²	1.5×10 ⁻²
4	6.8×10^{-2}	0.19	5.8×10^{-2}	9.4×10 ⁻³

4. Conclusions

The FeP glass with a nominal composition of $42P_2O_5$ -25Na₂O-5Al₂O₃-10CaF₂-18Fe₃O₄ (in mol %) was used for the immobilization of Re. The maximum retention of Re in the glass was as high as ~1.1 wt. % and its volatilization can be reduced to below ~50 %. The 7-day PCT normalized release of Re was less than 10^{-2} g/m², well below the current US DOE specification for high-level waste. Therefore, the FeP glass offers a significant potential for the immobilization of ⁹⁹Tc.

ACKNOWLEDGEMENTS

This work was supported by World Class University (WCU) (R31-30005) Program through the National Research Foundation of Korea (NRF) funded by the Ministry of Education, Science and Technology.

REFERENCES

[1] K. Schwochau, The Present Status of Technetium Chemistry, Radiochim. Acta, Vol.32, p.139, 1983.

[2] R. E. Wildung, K. M. Mcfadden and T. R. Garland, Technetium Sources and Behavior in the Environment, J. Environ. Qual., Vol.8, p.156, 1979.

[3] D. D. Keiser Jr, D. P. Abraham and J. W. Richardson Jr, Influence of Technetium on the Microstructure of a Stainless Steel-Zirconium Alloy, J. Nucl. Mater., Vol.277, p.333, 2000.
[4] M. Y. Khalil and W. B. White, Magnesium Titanate Spinel: a Ceramic Phase for Immobilization of Technetium-99 from Radioactive Wastes, J. Am. Ceram. Soc., Vol.66, p.197, 1983.

[5] D. Singh, V. R. Mandalika, S. J. Parulekar and A. S. Wagh, Magnesium Potassium Phosphate Ceramic for ⁹⁹Tc Immobilization, J. Nucl. Mater., Vol.348, p.272, 2006.

[6] D. E. Day, C. S. Ray, C. W. Kim, W. Huang, R. D. Leerssen, S. T. Reis, and D. Zhu, Iron Phosphate Glasses: An Alterative for Vitrification Certain Nuclear Wastes, Report for DOE Contract FG07-96ER45618, University of Missouri-Rolla, Rolla, MO 2002.

[7] W. L. Ebert, S. F. Wolf and J. K. Bate, The Release of Technetium from Defense Waste Processing Facility Glasses, Material Research Society Meeting, Boston, MA, Nov. 27 1995.